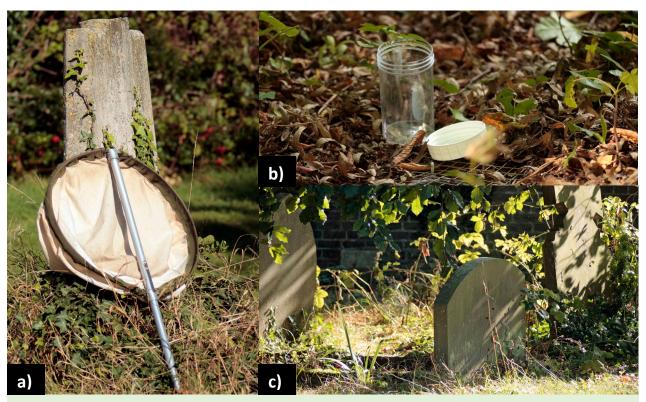
# **Comparing invertebrate sampling techniques in Mill Road Cemetery**

### Introduction

The importance of invertebrates in ecosystem processes puts their effective surveying in the frontlines of research. Biological inventories are needed in order to focus conservation efforts adequately (Barnosky et al. 2011) and understand the effects of urbanization (Faeth et al. 2011) on invertebrate populations. The study site was Mill Road Cemetery, Cambridge (Figure 1, 2) which is a neutral-calcareous grassland area and forms a mosaic heterogeneous habitat. Cemeteries are proven to be high biodiversity (Gilbert 1991, Laske 1994) patch areas within the urban matrix. Scientists have been sampling invertebrates for a centuries and therefore, there is a great variety of techniques known. They can be grouped in two ways; (1) it can be either qualitative or quantitative or (2) passive or active. In this study, four sampling methods (pitfall, sweep netting, tree beating and suction) were used to collect invertebrates. The null hypothesises were the following: (i) there is no difference between the sampling techniques and sample sizes; and (ii) there is no difference between sampling techniques and species composition. The ecological role of each taxa, environmental factors and the vertical stratification of invertebrates is also discussed.



**Figure 1.** Pictures of the different areas of Mill Road Cemetery, Cambridge. Amenity grass is kept at short length while False-oats are not cut back surrounding the graves (a, c). Pitfall trapping (b) was set up in the northern part of the cemetery which is only cut once or twice a year (Cambridge City Wildlife Site Survey 2005). Suction sampling, tree beating and sweep netting (a) was carried out in the regularly mowed area of the cemetery.

# **Study Site**

Mill Road Cemetery is 168 years old and is under two different management regimes. The main circular path is regularly mowed and is dominated by *Trifolium* spp., *Arrhenatherum elatius* and *Pheleum pratens*. The north end of the cemetery is only cut a few times per year. The density of trees (*Acer pseudoplatanus*, *Acer platanoides*, *Sambucus nigra*) and shrubs (*Rubus fructicosus* agg.) are higher in this area.

### **SITE NAME: Mill Road Cemetery**

Site code: G5.1 Grid ref: TL461582

Date of survey: 5/10/2016

#### **Habitat information**

Code Habitat type

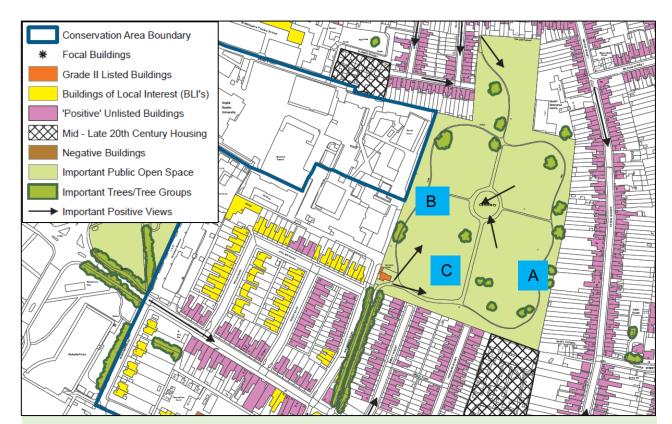
A33 Parkland, scattered trees; mixed
 B22 Grassland: neutral, semi-improved
 B32 Grassland: calcareous, semi-improved

J25 Boundaries, wall

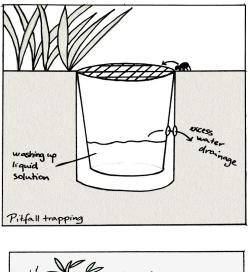
**Figure 2.** Phase 1 survey description of study site.

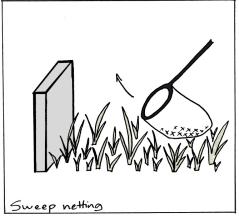
### **Methods**

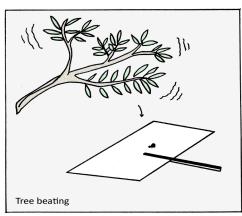
Four sampling methods were used to sample the cemetery. Pitfall trapping (Figure 2, 3) consisted of 27 plastic cups in the ground containing washing up solution. Its opening was on the ground level with a wire net on top for 3 days. Sweep netting and tree beating was carried out on the regularly managed area of the cemetery (location B). Insects were collected by pooting from the surface of the equipment. All samples were collected by two students but exact number of pairs is unknown. Samples were kept in ethanol solution. Suction sampling was carried out by an engine powered insect suction sampler (location C).



**Figure 2.** Map of the cemetery and its surroundings using Phase 1 habitat survey (original map by The Wildlife Trust for the Cambridge City Wildlife Site Survey 2005). The cemetery is a mosaic of different habitats (described on the right). Capital letters indicate the sampling areas.







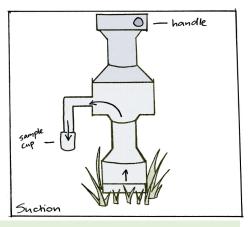
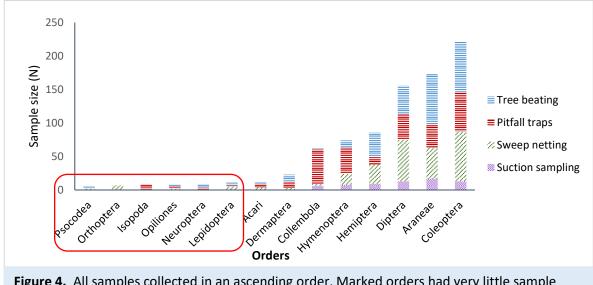


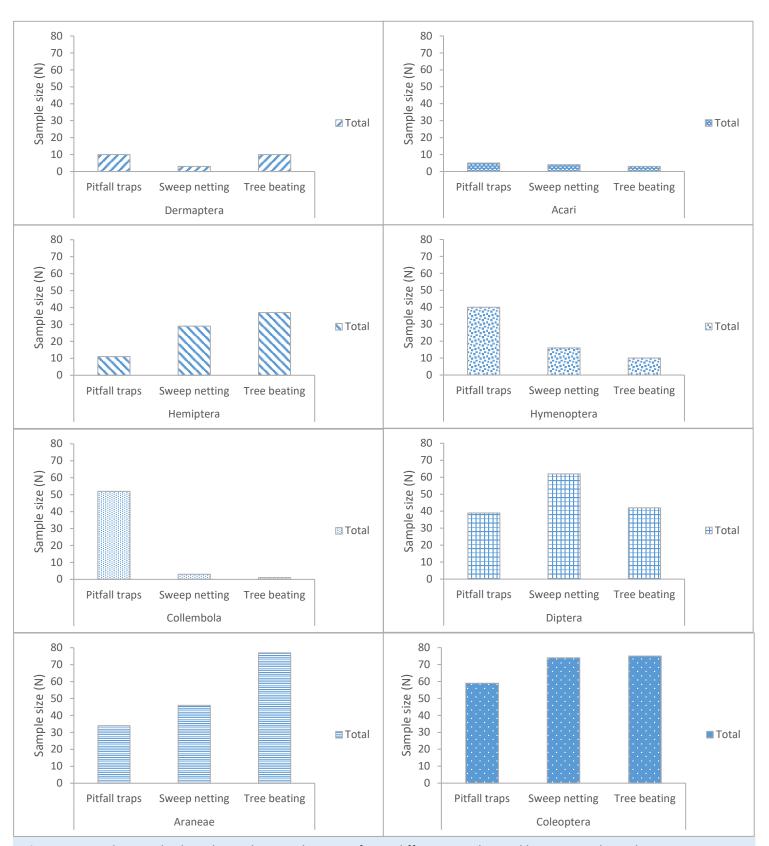
Figure 3. Four sampling methods used in this survey (illustrated by author).

### **Results**

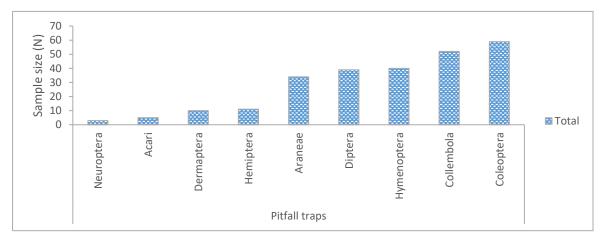
The total number of 860 individuals were identified to order level using compound microscopes. Suction samples were incomplete so the data was omitted (Appendix 1). Six orders contained very few individuals (Figure 4) which were also discarded. Therefore eight orders were used for the statistical analysis (Appendix 2). The three most abundant orders were the Coleoptera (n=221), Araneae (n=174) and the Diptera (n=156). For testing independence of observed frequencies, two way Chi square test was used for the nonparametric data. The test showed a highly significant difference between the sampling methods ( $X^2_3$ =156.630, N=742, p < 0.05) and their sample sizes (Figure 5).

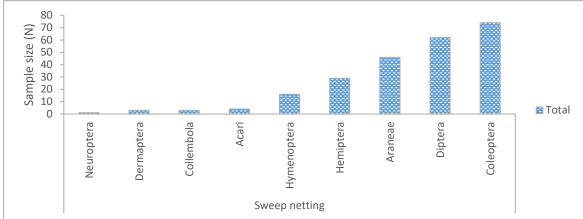


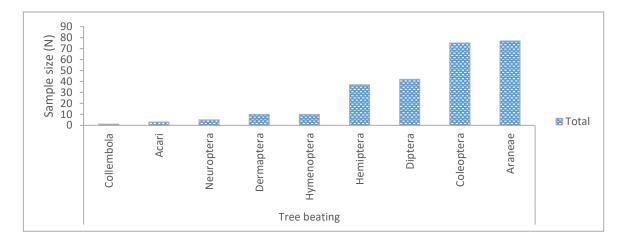
**Figure 4.** All samples collected in an ascending order. Marked orders had very little sample sizes and were discarded.

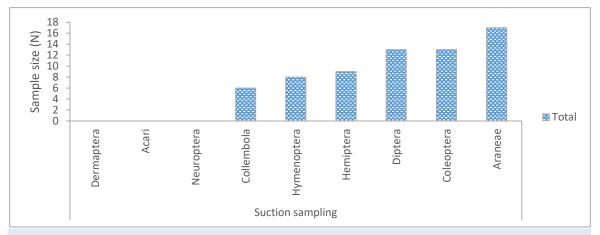


**Figure 5.** Sampling methods and sample sizes show significant difference within and between orders. The most significant differences are between the pitfall traps and the tree beating samples (Hemiptera, Hymenoptera, Collembola, Araneae). The difference can be explained by describing the different taxa's feeding behaviour, their habitats, dispersal strategies and development.









**Figure 5.** Sampling methods and sample sizes show significant difference within and between orders. The most significant differences are between the pitfall traps and the tree beating samples (Hemiptera, Hymenoptera, Collembola, Araneae). The difference can be explained by

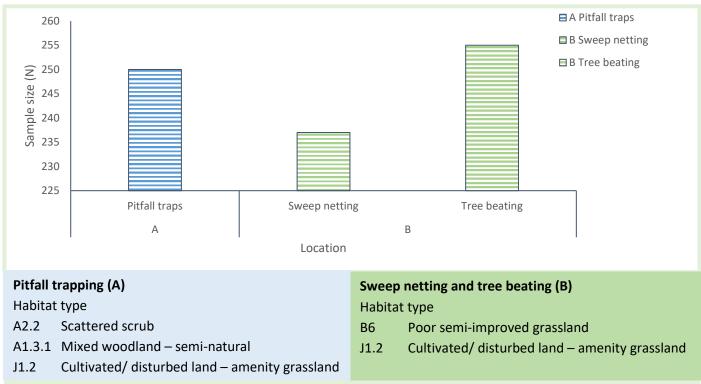
#### **Discussion – Taxa distribution**

The significant differences between sampling methods and sample composition shows that different invertebrate taxa occur in different (a) habitats and (b) on different vertical levels. Coleoptera, which is the largest of all insect orders, was the most abundant for pitfall trapping and sweep netting. Although beetles can fly, they mostly occupy the ground and the low vegetation which explains why their numbers greatly decreased when collected by tree beating. They have biting mouthparts and a tough elytra for physical protection which makes them a really successful species. Plant-feeders, wood-borers, scavengers, predators and parasites. Dung beetles and burying beetles are crucial links in the nitrogen cycle. Studies focusing on beetles usually use pitfall trapping for sampling (Thomas and Sleeper 1997). Hemiptera, compromised of Homoptera and Heteroptera suborders, is associated with roots, basal and aerial parts of vegetation which is consistent with our data, since we had more individuals from tree beating and sweep netting, than pitfall trapping. They are important agricultural pests, disease transmitters (Valente et al. 1998); mostly sapsuckers but rarely predators (Cohen 1990). 1650 species occur in the UK. Collembola has a world-wide distribution but they are susceptible to drying and therefore they are mostly found in soil or leaf litter. They have biting-mouthparts and their feeding varies from feeding on algae, pollen to decaying matter and nematodes (Brown 1954). Collembola the second most abundant taxa in the pitfall traps. As the traps were mostly located particularly underneath Acer pseudoplatanus and close to Rubus fructicosus the vegetation provided shade and a damp environment; ideal for springtails. Immature stages of Diptera can be highly abundant in the soil which explains the high number of captures in the pitfall traps. Araneae are mostly carnivorous (a few are phytophagous, Krantz and Lindquist, 1979), web-building species show pronounced vertical stratification with correlation to the vegetation (Duffey 1962, Enders 1974). This explains the lower numbers of pitfall trapping result compared to the tree beating (and suction) sampling method. Both Araneae and Acari belong to the class Arachnida. Their morphological similarities might have made it more difficult for the students to identify them correctly and that is why the Acari numbers are surprisingly low. **Hymenoptera**, containing over 100,000 known species, has a great variety of roles within ecosystems.

They carry out organic matter decomposition, soil fixation, predation, parasitism, herbivory, disease transmission and pollination.

Comparison between our sampling and the two management regimes (regularly mowed *versus* only cut once or twice per year, Figure 6) could have provided data for diversity measurement between the two areas. Because of non-uniform sampling, it cannot be tested statistically but can be hypothesized that assessing the ß-diversity between the two habitats might show difference in species richness and composition.

Our sampling methods can be compared used to compare vertical heterogeneity as opposed to horizontal one. There are numerous techniques to sample different habitats (Table 1). Many studies focus on the comparison of the effectiveness of different sampling methods used for different taxa (Corti et al. 2013, Basset et al. 1997, Hurley et al. 2015, Bouget et al. 2009, Alonso and Camargo 2010, Doxon et al. 2013). In order to model an ecosystem it is crucial to have uniform sampling effort, be qualitative and also quantitative. However, when considering the different lifestyles of animals, it is impractical to use only one method as our study has proven.



**Figure 6.** Sampling was carried out in two different habitats (Figure 2) which will affect the invertebrate composition. Statistically cannot be compared because of non-uniform sampling effort (27 pitfall traps *versus* unknown number of individual samples of sweep netting and tree beating).

There is much debate about the significant result as only qualitative sampling methods were used for the statistical analysis. The distribution of sampling is another major issue when comparing the samples. Sweep netting and tree beating was not clearly defined, students also sampled shrubs. The lack of surveying plant coverage and composition is also decreasing the effectiveness of this survey. The mishandling the suction samples led to omitting all the data and reduced the discussion of four methods to three.

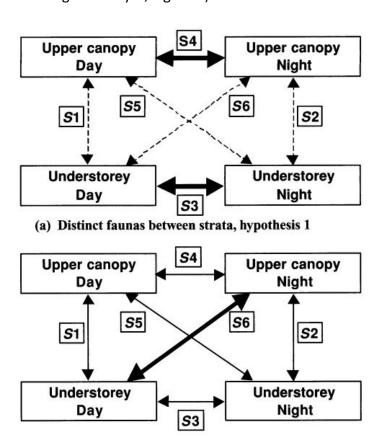
**Table 1.** Different habitats require different methods. Even within one habitat several complementary methods will be able to assess relatively precisely the invertebrate composition. Adapted from Samways et al. 2010.

Target habitat	Methods available		
Forest canopy	Direct access via cranes, platforms, ropes Fogging/ misting using insecticide Direct foliage samples		
Low canopy, shrubs	Sweep netting Beating Selective individual collection		
Low vegetation	Sweep netting Suction sampling Direct examination		
Ground/ surface	Pitfall Soil and litter sifting Tullgren-Berlese funnel extraction Winkler bag extraction		
Flying insects	Malaise traps Window/ interception traps Suction traps Light traps Pheromone/bait traps Pan traps/water traps		
Aquatic insects	Waterside vegetation Dip netting Surber sampling Benthic grabs		

#### Discussion - Vertical distribution

The significant difference between the three sampling methods shows that for comprehensive invertebrate survey requires several sampling techniques, not just one (Table 1). Horizontal heterogeneity can be surveyed by a species-area model which requires precise distance data. The complexity of vertical heterogeneity depends on habitat structure (desert system could be less complex than a rainforest). As the vegetation strata changes, so does the arthropod assemblages (Smith 1973). Ground-to-canopy transects

are crucial in order to produce eloquent data set for habitat modelling. As there is such a great variety of sampling methods, one must consider first of all the ecological question of the study and consider both (a) long-term and (b) short-term variations between the vertical stratas. A long-term example is the change of distribution throughout invertebrate development stages. Short-term example is the arthropod diel activity studied in rainforests (Basset et al. 2001, Figure 7) by using tree beating, flight-interception traps and sticky traps. Considering short-term variation, environmental variables (EV, Figure 8.) will have an effect on invertebrates which can also be (i) long-term or (ii) short-term. Short-term EV might be solar radiation, fluctuation of relative humidity, air temperature, wind speed, water condensation, canopy openness (Barker 1996, Parker 1996). Invertebrate assemblages will also differ among guilds (e.g. scavengers, deadwood eaters, herbivores, predators, pollinators, parasitoids, fungal feeders etc) (Basset 2003). Most importantly, resource quality and quantity has a strong influence on trophic structures (Price 1992) which leads to vertical strata based on different feeding guilds (such as more herbivores and their predators in the canopy and more decomposers, fungivores in the ground layer, Figure 9).



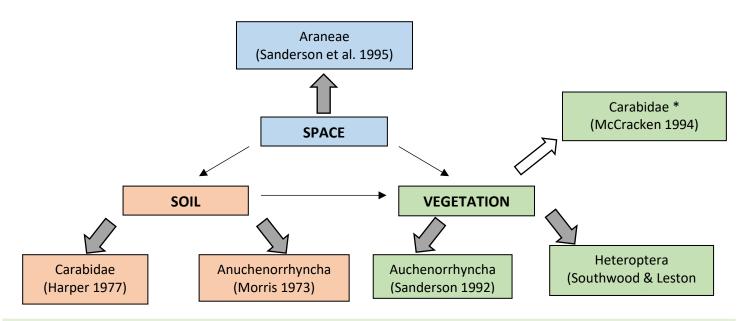
**Figure 7.**Hypothesis by Basset et al. (2001) of faunal correlation with invertebrate diel activity in a rainforest. S1 to S6 are coefficients of similarity between time of the day and fauna. Their data identified significant differences between day and night abundances.

■ Predator

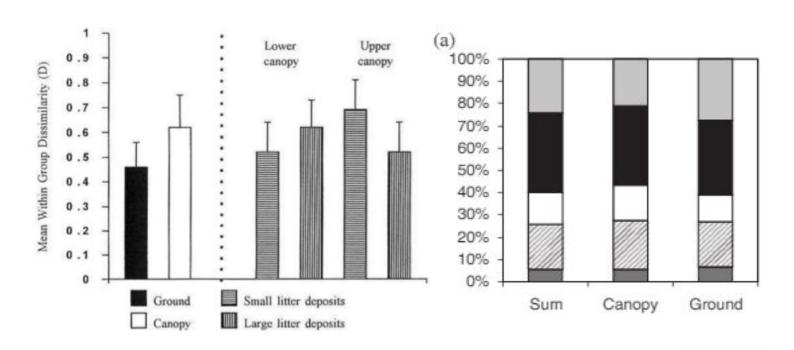
Herbivore

☐ Xylophage
☐ Fungivore

■ Saprophage



**Figure 8.** Environmental variables affecting different invertebrate taxa. Kowarik et al. (2016) found that plant  $\alpha$ -diversity negatively relates to increasing tree layer and presence of *Hedera helix* in a 53 ha German cemetery. (Illustration adapted from Kowarik et al. 2016)



**Figure 9.** Collembola vertical stratification in rainforest (left) (Rodgers and Kitching 1998) showing dissimilarity between samples from the ground and the canopy and even within the lower and upper canopy in an Australian rainforest. Beetles also show segregation according to both of their feeding guilds' and body size (on the right; proportional representation by species) (Grimbacher and Stork 2007)

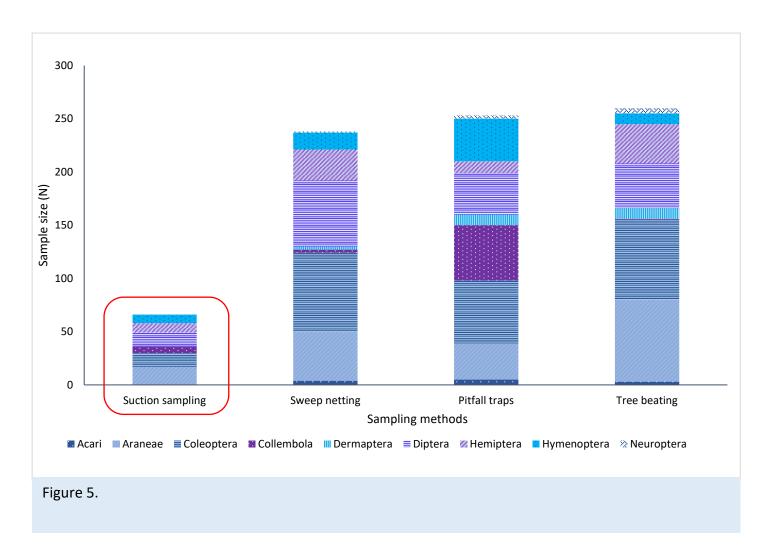
# Conclusion

Invertebrate sampling in Mill Road Cemetery, Cambridge showed the importance of planning a survey and using multiple methods to assess the species composition of different habitats and strata of heterogeneity. The significant difference between sampling methods identified how different invertebrate taxa occupy different horizontal and vertical strata. Applying this knowledge can help healthcare professionals to track and stop the spread of diseases (e.g. dengue fever, Romero-vivas et al. 2002) or use invertebrates for forensic science (e.g. Sharanowski et al. 2008). As invertebrates are also key prey resource (Vonshak et at. 2009). IFor example, invertebrate abundance has a strong relation with chick survival (Palmer et al. 2001, Mitchell and Riegert 1994) therefore their survey is also crucial to bird conservation plans. From a global respective, reserving ecological networks require \$3-11 billion per year between 2000-2030 (James et al. 2001, Pimm et al. 2001). Surveying and understanding invertebrate assemblages is in the frontlines of research and there is a great need for continuous development and testing (Meir and Possingham 2004).

**Total Word Count: 1593** 

# **Appendix**

		Grand		
Orders	Sweep	Pitfall	Tree	Total
	netting	trapping	beating	Total
Acari	4	5	3	12
Araneae	46	34	77	157
Coleoptera	74	59	75	208
Collembola	3	52	1	56
Dermaptera	3	10	10	23
Diptera	62	39	42	143
Hemiptera	29	11	37	77
Hymenoptera	16	40	10	66
Grand Total	237	250	255	742



#### References

Alonso, Á. and Camargo, J.A., 2010. Comparison of two methods for estimating the abundance, diversity and habitat preference of fluvial macroinvertebrates in contrasting habitats. *Limnologica-Ecology and Management of Inland Waters*, 40(1), pp.23-29.

**Bardgett,** R.D. and Wardle, D.A., 2003. Herbivore-mediated linkages between aboveground and belowground communities. *Ecology*, *84*(9), pp.2258-2268.

Barnosky, A.D., Matzke, N., Tomiya, S., Wogan, G.O., Swartz, B., Quental, T.B., Marshall, C., McGuire, J.L., Lindsey, E.L., Maguire, K.C. and Mersey, B. (2011) Has the Earth/'s sixth mass extinction already arrived? *Nature*, Vol 471(7336), pp.51-57

Basset, Y., 2003. *Arthropods of tropical forests: spatio-temporal dynamics and resource use in the canopy.* Cambridge University Press.

Basset, Y., ABERLENC, H.P., Barrios, H., Curletti, G., BÉRENGER, J.M., VESCO, J.P., Causse, P., Haug, A., HENNION, A.S., LESOBRE, L. and MARQUÈS, F., 2001. Stratification and diel activity of arthropods in a lowland rainforest in Gabon. *Biological Journal of the Linnean Society*, 72(4), pp.585-607.

Basset, Y., Springate, N.D., Aberlenc, H.P. and Delvare, G., 1997. A review of methods for sampling arthropods in tree canopies. *Canopy arthropods. Chapman & Hall, London*, pp.27-52.

Bouget, C., Brustel, H., Brin, A. and Valladares, L., 2009. Evaluation of window flight traps for effectiveness at monitoring dead wood-associated beetles: the effect of ethanol lure under contrasting environmental conditions. *Agricultural and Forest Entomology*, 11(2), pp.143-152.

Brown, W.L., 1954. Collembola feeding upon nematodes. *Ecology*, 35(3), pp.421-421.

Cohen, A.C., 1990. Feeding adaptations of some predaceous Hemiptera. *Annals of the Entomological Society of America*, 83(6), pp.1215-1223.

Corti, R., Larned, S.T. and Datry, T., 2013. A comparison of pitfall-trap and quadrat methods for sampling ground-dwelling invertebrates in dry riverbeds. *Hydrobiologia*, 717(1), pp.13-26.

Doxon, E.D., Davis, C.A. and Fuhlendorf, S.D., 2011. Comparison of two methods for sampling invertebrates: vacuum and sweep-net sampling. *Journal of Field Ornithology*, 82(1), pp.60-67.

Duffey, E., 1962. A population study of spiders in limestone grassland. *The Journal of Animal Ecology*, pp.571-599.

Enders, F., 1974. Vertical Stratification in Orb-Web Spiders (Araneidae, Araneae) and a Consideration of Other Methods of Coexistence. *Ecology*, 55(2), pp.317-328.

Faeth, S.H., Bang, C. and Saari, S. (2011) Urban biodiversity: patterns and mechanisms. *Annals of the New York Academy of Sciences*, Vol 1223(1), pp. 69-81

Grimbacher, P.S. and Stork, N.E., 2007. Vertical stratification of feeding guilds and body size in beetle assemblages from an Australian tropical rainforest. *Austral Ecology*, *32*(1), pp.77-85.

Harper, J.L. (1977) Population Biology of Plants. Academic Press: London

Hurley, B.P., Garnas, J. and Cooperband, M.F., 2015. Assessing trap and lure effectiveness for the monitoring of Sirex noctilio. *Agricultural and Forest Entomology*, *17*(1), pp.64-70.

Kearsley, M.J. and Whitham, T.G., 1989. Developmental changes in resistance to herbivory: implications for individuals and populations. *Ecology*, 70(2), pp.422-434.

Krantz, G.W. and Lindquist, E.E., 1979. Evolution of phytophagous mites (Acari). *Annual Review of Entomology*, 24(1), pp.121-158.

McCraken, D.I. (1994) A fuzzy classification of moorland ground beetle (Coleoptera: Carabidae) and plant communities. *Pedobiologia*, Vol 38, pp. 12-27

Meir, E., Andelman, S. and Possingham, H.P., 2004. Does conservation planning matter in a dynamic and uncertain world?. *Ecology Letters*, 7(8), pp.615-622.

Mitchell, G.J. and Riegert, P.W., 1994. Sharp-tailed grouse, Tympanuchus phasianellus, and grasshoppers: food is when you find it. *Canadian field-naturalist*. *Ottawa ON*, *108*(3), pp.288-291.

Morris, M.G. (1973) The effects of seasonal grazing on the Heteroptera and Auchenorrhyncha (Hemiptera) of chalk grassland. J. Appl. Ecol. 10, 761-80.

Nyman, T., Widmer, A. and Roininen, H., 2000. Evolution of gall morphology and host-plant relationships in willow-feeding sawflies (Hymenoptera: Tenthredinidae). Evolution, 54(2), pp.526-533.

Palmer, W.E., Lane, M.W. and Bromley, P.T., 2001. Human-imprinted northern bobwhite chicks and indexing arthropod foods in habitat patches. The Journal of wildlife management, pp.861-870.

Rodgers, D.J. and Kitching, R.L., 1998. Vertical stratification of rainforest collembolan (Collembola: Insecta) assemblages: description of ecological patterns and hypotheses concerning their generation. Ecography, 21(4), pp.392-400.

Rodgers, D.J. and Kitching, R.L., 1998. Vertical stratification of rainforest collembolan (Collembola: Insecta) assemblages: description of ecological patterns and hypotheses concerning their generation. Ecography, 21(4), pp.392-400

Romero-Vivas, C.M., Wheeler, J.G. and Falconar, A.K., 2002. An inexpensive intervention for the control of larval Aedes aegypti assessed by an improved method of surveillance and analysis. Journal of the American *Mosquito Control Association, 18*(1), pp.40-46.

Samways, M.J., McGeoch, M.A and New, T.R. 2010. Insect conservation: a handbook of approaches and methods. Oxford: Oxford University Press, pp. 90-92

Sanderson, R.A, Rushton, S.P., Cherill, A.J. and Byrne J.P. (1995) Soil, vegetation and Space: an analysis of their effects on invertebrate communities of a moorland in North-east England. Journal of Applied Ecology, Vol 32 (3), pp. 506-518

Sanderson, R.A. (1992) Diversity and evenness of Hemiptera communities on naturally vegetated derelict land in NW England. Ecography, Vol 15, pp. 154-160

Sharanowski, B.J., Walker, E.G. and Anderson, G.S., 2008. Insect succession and decomposition patterns on shaded and sunlit carrion in Saskatchewan in three different seasons. Forensic Science International, 179(2), pp.219-240.

Smith, A.P., 1973. Stratification of temperature and tropical forests. *American Naturalist*, pp.671-683.

Southwood, T.R.E. and Leston, D. (1959) Land and Water Bugs of the Birtish Isles. Frederick Warne & Co: London

Stork, N.E., 1988. Insect diversity: facts, fiction and speculation. Biological journal of the Linnean Society, 35(4), pp.321-337.

Thomas, D.B. and Sleeper, E.L., 1977. The use of pit-fall traps for estimating the abundance of arthropods, with special reference to the Tenebrionidae (Coleoptera). Annals of the Entomological Society of America, 70(2), pp.242-248.

Valente, V.C., Valente, S.A., Noireau, F., Carrasco, H.J. and Miles, M.A., 1998. Chagas disease in the Amazon Basin: association of Panstrongylus geniculatus (Hemiptera: Reduviidae) with domestic pigs. *Journal of Medical Entomology*, 35(2), pp.99-103.

Vonshak, M., Dayan, T. and Kronfeld-Schor, N., 2009. Arthropods as a prey resource: patterns of diel, seasonal, and spatial availability. *Journal of Arid Environments*, 73(4), pp.458-462.